

Mangrove vegetation in Amazonia: a review of studies from the coast of Pará and Maranhão States, north Brazil

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ABSTRACT

The present study is a compilation of the literature about vegetation of mangrove forest of the north coast of Brazil. It synthesizes the knowledge about this important ecosystem and lists the currently available literature. The study focuses on the coast of Pará and Maranhão states, which are covered by a continuous belt of mangroves. The mangrove flora comprises six mangrove tree species and several associated species. Mangrove tree height and stem diameter vary as a function of abiotic local stand parameters. Seasonal variation in rainfall and salinity affect the species' phenology and litter fall. Local population use products derived from mangrove plants for different purposes (e.g. fuel; medicinal; rural construction). The increase in the coastal population has given rise to conflicts, which impact on mangrove forest.

KEY WORDS: Amazonian mangroves, phenology, litter fall, herbivory, anthropic impacts

Vegetação de manguezais na Amazônia: uma revisão dos estudos da costa dos Estados Pará e Maranhão, norte do Brasil

RESUMO

O presente estudo apresenta uma compilação da literatura sobre a vegetação dos manguezais da costa norte do Brasil, apresentando uma síntese do conhecimento e listando a literatura disponível. O estudo se concentra na costa dos estados do Pará e Maranhão que formam um cinturão contínuo de manguezais. Foram contabilizadas seis espécies arbóreas exclusivas de mangue e várias outras associadas. A altura e o diâmetro das árvores de mangue variam em função de parâmetros abióticos locais. As variações sazonais do regime de chuvas e da salinidade afetam a fenologia das espécies e a produção de serapilheira. A população costeira utiliza a flora do manguezal para diferentes fins (ex: combustível, medicinal, construção rural). O aumento da ocupação costeira inicia um processo de impacto para as florestas de mangue e a disponibilidade de seus recursos.

PALAVRAS CHAVE: Manguezais Amazônicos, fenologia, serapilheira, herbivoria, impacto antrópico

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INTRODUCTION

Mangroves are coastal forests of sheltered tropical and subtropical coastlines. They feature trees with a capacity to survive in a saline or brackish environment under tidal influence (Tomlinson, 1986). Mangroves offer a variety of economically important resources exploited by the coastal human population, and are important as fish nursery (Macnae, 1968; Sasekumar *et al.*, 1992) as well as for coastal protection (Ewel *et al.*, 1998). Their primary production is fundamental for estuarine and coastal food chains (Alongi *et al.*, 1989). Nevertheless, the mangrove ecosystem has been destroyed by construction of shrimp farms, by deforestation, wood cutting, overfishing, pollution and urban development (Spalding *et al.*, 1997). In Brazil, mangroves occur from the Amapá state in the North to Santa Catarina state in the South. The north coast has an area of approximately 8000 km² of mangroves (Souza-Filho, 2005). Along the coast the states of Pará and Maranhão, mangroves form a continuous belt of about 700000 ha, that is, nearly 85% of all Brazilian mangroves (Herz, 1991; Lacerda, 1999). In recent years, integrated research into the coastal ecosystems of this area has been initiated (Berger *et al.*, 1999; Prost *et al.*, 2001) and has provided most of the currently available knowledge on mangrove vegetation of the region. However, no summary of this information is available. This review tries to fill this gap by presenting a synthesis of studies on mangrove vegetation of the coasts of Pará and Maranhão states.

DATA COLLECTION

The literature analysed includes papers published in peer reviewed journals, books, conference proceedings, internal reports of university programs as well as graduate, master and doctorate theses. Examinations of mangrove specimens from the collection of the Herbarium João Murça Pires at the Museu Paraense Emílio Goeldi at Belem (MG) and personal observations complete the information.

STUDY AREA

We concentrate on the coast of the north-Brazilian states of Pará and Maranhão, extending over 650 km from Marajó Bay (Pará State) to the estuary of river Preguiças (Maranhão State) between 0°13'45"S, 48°33'19"W and 2°29'32"S, 43°27'44"W (Figure 1). In this region, mangroves can extend more than 40 km landward following the course of numerous small estuaries and bays (Souza-Filho, 2005). The area is characterised as macrotidal; tidal amplitude varies between 4 m and 7.5 m (Souza-Filho, 2005). The tidal regime is semi-diurnal (DHN, 2004). The climate is tropical, warm and humid. The annual mean precipitation is approximately 2300 mm (Fisch *et al.*, 1998). The mean air temperature is about 26°C. The region presents a well defined dry season

with mean monthly precipitation less than 50 mm, lasting from July to December (Fisch *et al.*, 1998).

TERMINOLOGY

The present study considers three classes of pore water salinity, according to Santos *et al.* (1997): brackish, saline and hypersaline, with salinities of <30; 30-60 and >60, respectively.

RESULTS AND DISCUSSION

FLORISTIC COMPOSITION AND SPECIES DISTRIBUTION

Major elements of the mangrove forest flora (*sensu* Tomlinson, 1986)

Rhizophora mangle L. (Rhizophoraceae) is the most widely distributed mangrove tree species (Prance *et al.*, 1975), and is dominant in estuaries more exposed to the ocean (Almeida, 1996a). A second, less widely distributed *Rhizophora* species is *Rhizophora racemosa* G.F.W. Meyer, which occurs in Marajó Bay (Almeida, 1996a). Further south, there is only one other report of *R. racemosa*, from the Preguiças estuary, Maranhão (Santos, 1986). The third species of the genus, *Rhizophora harrisonii* Leechman, is less frequently reported than the other two species (Figure 2). Only in the Preguiças estuary it is recorded as dominant mangrove tree (Santos, 1986). According to Breteler (1969) and Lima *et al.* (2005), the salt tolerance of the *Rhizophora* species increases from less tolerant *R. racemosa* through *R. harrisonii* to *R. mangle*. Low salt tolerance might explain the restriction of *R. racemosa* and *R. harrisonii* to the Marajó bay region (where pore water salinity is about 5 - 12, Menezes, unpubl. data; Figure 2) and the Preguiças estuary. However, no detailed investigations have been carried out into the salt tolerance of these species.

Avicennia germinans (L.) L. is the commonest species of the family Avicenniaceae in our area (Figure 3). According to Luz *et al.* (2000), *A. germinans* is more frequent in elevated, less inundated areas and under more saline conditions.

Avicennia schaueriana Stapf and Leechman ex Moldenke has been recorded at several locations along the coast (Figure 3), principally near sandy beaches (Amaral *et al.*, 2001; Santos, 2005). This species has nowhere been reported as a major constituent of the mangrove forest in our study area.

Laguncularia racemosa (L.) C.F. Gaertn. (Combretaceae) occurs along the entire coast in saline as well as in brackish water mangroves (Figure 3). This species mostly occupies forest edges, large gaps, riparian sites, and other open areas. There are no records of *L. racemosa* dominated forests.

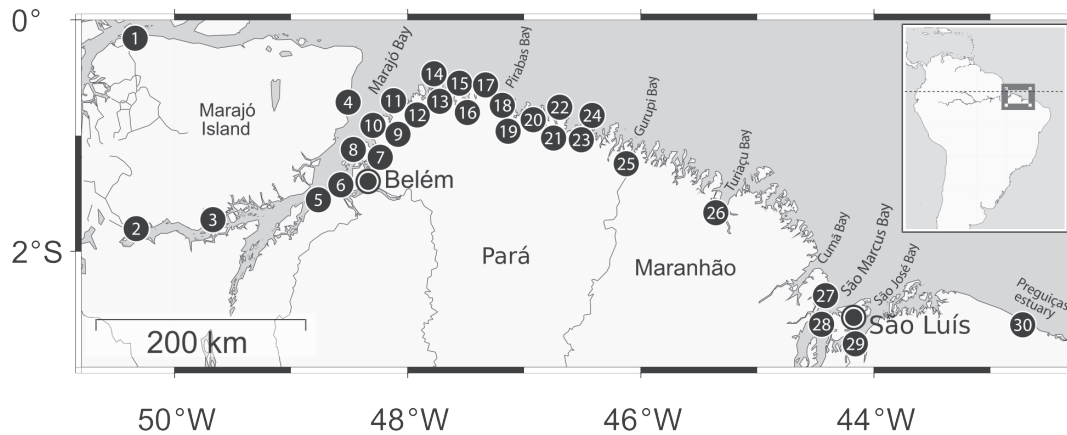


Figure 1 - Study area on the North Coast of Brazil, states of Pará (with the Marajó Island) and Maranhão. Localities with informations about mangrove forests are indicated by numbers (1: Afuá; 2: Breves; 3: São Sebastião da Boa Vista; 4: Soure, Salvaterra; 5: Barcarena; 6: Ilha das Onças; 7: Ananindeua, Marituba and Benevides; 8: Ilha do Mosqueiro, Santa Bárbara; 9: Vigia; 10: Colares; 11: São Caetano de Odivelas; 12: Curuçá; 13: Mapanim; 14: Marudá, Praia do Crispim; 15: Ilha de Algodoal; 16: Maracanã; 17: Salinas (Salinópolis); 18: São João de Pirabas; 19: Primavera; 20: Quatipuru; 21: Bragança; 22: Ilha de Canelas; 23: Augusto Corrêa; 24: Urumajó; 25: Viseu; 26: Turiaçu; 27: Alcântara; 28: Baía de São José; 29: Baía de São Marcus, Parnaúçu e Rio Mearim 30: Rio Preguiças). For more informations about the geographical position of each site see Annexes.

Table 1 - Minor elements of mangrove forest and species associated to mangrove flora. Bs: brackish mangrove; hs = saline mangrove.

Family	Species	Sal.		Reference	Family	Species	Sal.		Reference	
		br	hs				br	hs		
Aizoaceae	<i>Sesuvium portulacastrum</i>		♦	6; pers. obs.	Dennstaedtiaceae	<i>Pteridium aquilinum</i>		♦	1	
Amaranthaceae	<i>Alternanthera</i> sp.		♦	1	Euphorbiaceae	<i>Alchornea brevistyla</i>		♦	1	
	<i>Blutaparon</i> sp.		♦	6	Fabaceae	<i>Desmodium canum</i>		♦	1	
Amaryllidaceae	<i>Crinum</i> sp.		♦	10; 2; 11; 20		<i>Machaerium lunatum</i>		♦	9	
Annonaceae	<i>Annona palustris</i>		♦	15		<i>Pterocarpus</i> spp.		♦	12	
Apocynaceae	<i>Mesechites trifidus</i>		♦	1		<i>Macherium lunatum</i>		♦	13	
	<i>Echites valenzuelanus</i>		♦	1		<i>Pterocarpus rohrii</i>		♦	15	
	<i>Rhabdadenia biflora</i>		♦	♦	pers. obs.		<i>Drepanocarpus lunatus</i>		♦	15
	<i>Forsteronia</i> sp.		♦			<i>Inga</i> sp.		♦	15	
	<i>Mandevilla</i> sp.		♦	18; 19		<i>Muelleria frutescens</i>		♦		
Araceae	<i>Montrichardia arborescens</i>		♦	1; 2	Lythraceae	<i>Crenea maritima</i>		♦	pers. obs.	
Arecaceae	<i>Euterpe oleracea</i>		♦	13	Malvaceae	<i>Pavonia spicata</i>		♦		
	<i>Mauritia flexuosa</i>		♦	13		<i>Hibiscus</i> sp.		♦	6	
	<i>Astrocaryum vulgare</i>		♦	18		<i>Hibiscus tiliaceus</i>		♦	19; 18	
Bataceae	<i>Batis maritima</i>		♦	6; pers. obs.	Melastomataceae	<i>Mouriri angulicosta</i>		♦	1	
Bombacaceae	<i>Bombax</i> sp.		♦	12		<i>Miconia</i> sp.		♦	12	
	<i>Bombax aquaticum</i> (= <i>Pachira aquatica</i>)		♦	13; 15	Orchidaceae	<i>Epidendrum ciliare</i>		♦	pers. obs.	
Cecropiaceae	<i>Cecropia</i> sp.		♦	6		<i>Epidendrum</i> sp.		♦	pers. obs.	
Clusiaceae	<i>Symphonia globulifera</i>		♦	13	Poaceae	<i>Sporobolus virginicus</i>		♦	2; 6	
Combretaceae	<i>Conocarpus erectus</i>		♦	3; 4; 5; 6; 7; 8; 17		<i>Spartina alterniflora</i>		♦	♦	2; 6
Costaceae	<i>Costus arabicus</i>		♦	1		<i>Spartina brasiliensis</i>		♦	14; 19	
Cyperaceae	<i>Cyperus giganteus</i>		♦	1		<i>Cynodon</i> sp.		♦	1	
	<i>Bulbostylis paraensis</i>		♦	1		<i>Paspalum</i> sp.		♦	6	
	<i>Cyperus distans</i>		♦	1					2; 6; 5; 14;	
	<i>Cyperus ligularis</i>		♦	1	Pteridaceae	<i>Acrostichum aureum</i>		♦	16; 18; 19;	
	<i>Fimbristylis</i> sp.		♦	6					18	
	<i>Eleocharis caribaea</i> sp.		♦	6; pers. obs.	Rhizophoraceae	<i>Cassipourea guianensis</i>		♦	15	
	<i>Eleocharis geniculata</i>			19						
	<i>Eleocharis intersita</i>		♦	6					References: 1 Tourinho (1998); 2 Menezes et al. (2003); 3 Medina et al. (2001); 4 Medina et al. (2000); 5 Lisboa et al. (1993); 6 Rebelo-Mochel. (1997); 7 Rebelo-Mochel et al. (2001); 8 Carreira et al. (2002); 9 Adams and Berger (2002); 10 Berger et al. (2006); 11 Matni (2003); 12 Gama et al. (1996); 13 Almeida (1996a); 14 Prost et al. (2001); 15 Ferreira (1989); 16 Sales (2000) 17 Santos et al. (2003); 18 Almeida et al. (2002); 19 Damasio (1980b); 20 Matni et al. (2006)	

Table 2 - Mangrove forest structure. Height: mean height; dbh: mean stem diameter; BA Rhz: *Rhizophora* Basal Area; BA Av: *Avicennia* Basal Area; BA Lg: *Laguncularia* Basal Area; Ag %: relative density of *Avicennia germinans*; As %: relative density of *Avicennia schaueriana*; Rm %: relative density of *Rhizophora mangle*; Rr %: relative density of *Rhizophora racemosa*; Rh %: relative density of *Rhizophora harrisonii*; Lg %: relative density of *L. racemosa*; Rr %: relative density of *Laguncularia racemosa*. Locality: local of each study. For more informations of localities position see Figure 1 and Annexes.

Family	Species	Sal.		Reference
		br	hs	
Aizoaceae	<i>Sesuvium portulacastrum</i>	●		5; pers. obs.
Amaranthaceae	<i>Alternanthera</i> sp.	●		1
	<i>Blutaparon</i> sp.		●	5
Amaryllidaceae	<i>Crinum</i> sp.	●		9; 2; 10;19
Annonaceae	<i>Annona palustris</i>	●		14
Apocynaceae	<i>Mesechites trifidus</i>	●		1
	<i>Echites valenzuelanus</i>	●		1
	<i>Rhabdadenia biflora</i>	●	●	pers. obs.
	<i>Forsteronia</i> sp.	●		
	<i>Mandevilla</i> sp.	●		17; 18
Araceae	<i>Montrichardia arborescens</i>	●		1; 2
Arecaceae	<i>Euterpe oleracea</i>	●		12
	<i>Mauritia flexuosa</i>	●		12
	<i>Astrocaryum vulgare</i>	●		17
Bataceae	<i>Batis maritima</i>		●	6; pers. obs.
Bombacaceae	<i>Bombax</i> sp.	●		11
	<i>Bombax aquaticum</i> (= <i>Pachira aquatica</i>)	●		12; 14
Cecropiaceae	<i>Cecropia</i> sp.	●		6
Clusiaceae	<i>Symphonia globulifera</i>	●		12
Combretaceae	<i>Conocarpus erectus</i>		●	3; 4; 5; 6; 7; 8; 16
Costaceae	<i>Costus arabicus</i>	●		1
Cyperaceae	<i>Cyperus giganteus</i>	●		1
	<i>Bulbostylis paraensis</i>	●		1
	<i>Cyperus distans</i>	●		1
	<i>Cyperus ligularis</i>	●		1
	<i>Fimbristylis</i> sp.		●	5

Family	Species	Sal.		Reference
		br	hs	
	<i>Eleocharis caribaea</i> *		●	5; pers. obs.
	<i>Eleocharis geniculata</i> *		●	18
	<i>Eleocharis intersita</i>		●	5
Dennstaedtiaceae	<i>Pteridium aquilinum</i>	●		1
Euphorbiaceae	<i>Alchornea brevistyla</i>	●		1
Fabaceae	<i>Desmodium canum</i>	●		1
	<i>Machaerium lunatum</i>	●		8;12
	<i>Pterocarpus</i> spp.	●		11
	<i>Pterocarpus rohrii</i>	●		14
	<i>Drepanocarpus lunatus</i>	●		14
	<i>Inga</i> sp.	●		14
	<i>Muelleria frutescens</i>	●		
Lythraceae	<i>Crenea maritima</i>	●		pers. obs.
Malvaceae	<i>Pavonia spicata</i>	●		
	<i>Hibiscus</i> sp.	●		5
	<i>Hibiscus tiliaceus</i>	●		18; 17
Melastomataceae	<i>Mouriri angulicosta</i>	●		1
	<i>Miconia</i> sp.	●		11
Orchidaceae	<i>Epidendrum ciliare</i>	●		pers. obs.
	<i>Epidendrum</i> sp.	●		pers. obs.
Poaceae	<i>Sporobulus virginicus</i>		●	2; 5
	<i>Spartina alterniflora</i>	●	●	2; 5
	<i>Spartina brasiliensis</i>		●	13; 18
	<i>Cynodon</i> sp.	●		1
	<i>Paspalum</i> sp.		●	5
Pteridaceae	<i>Acrostichum aureum</i>	●		2; 5; 4; 13; 15; 17; 18; 17
Rhizophoraceae	<i>Cassipourea guianensis</i>	●		14

References: 1 Tourinho (1998); 2 Menezes *et al.* (2003); 3 Medina *et al.* (2001); 4 Lisboa *et al.* (1993); 5 Rebelo-Mochel (1997); 6 Rebelo-Mochel *et al.* (2001); 7 Carreira *et al.* (2002); 8 Adams and Berger (2002); 9 Berger *et al.* (2006); 10 Matni (2003); 11 Gama *et al.* (1996); 12 Almeida (1996a); 13 Prost *et al.* (2001); 14 Ferreira (1989); 15 Sales (2000) 16 Santos *et al.* (2003); 17 Almeida *et al.* (2002); 18 Damasio (1980b); 19 Matni *et al.* (2006); * synonymous.

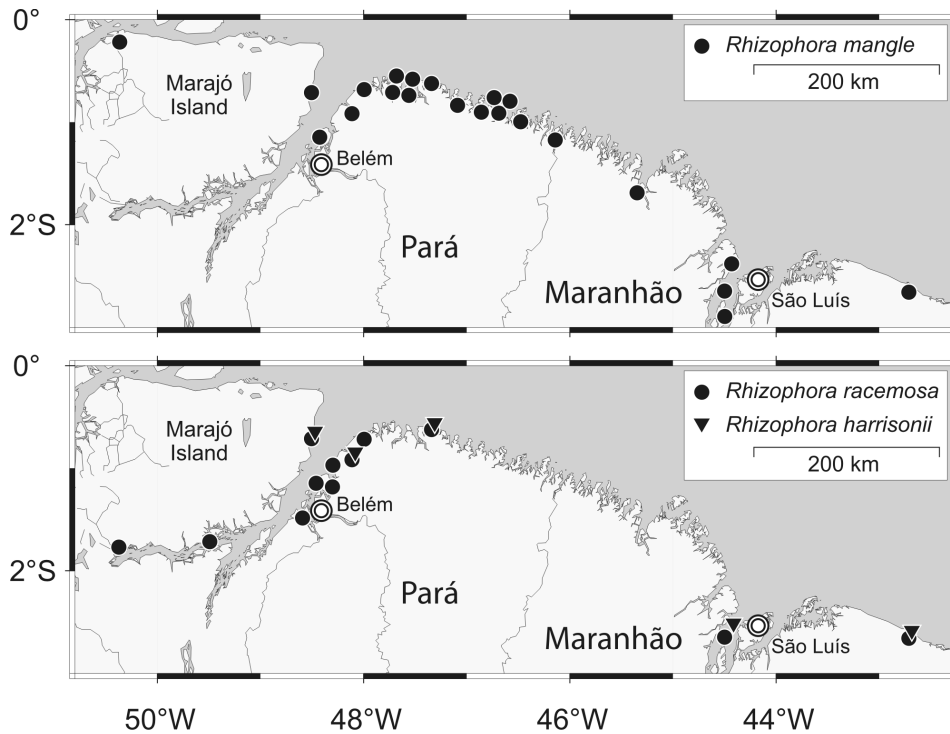


Figure 2 - Distribution of *Rhizophora* species in the study area. For details about the localities name and position, see Annex 1.

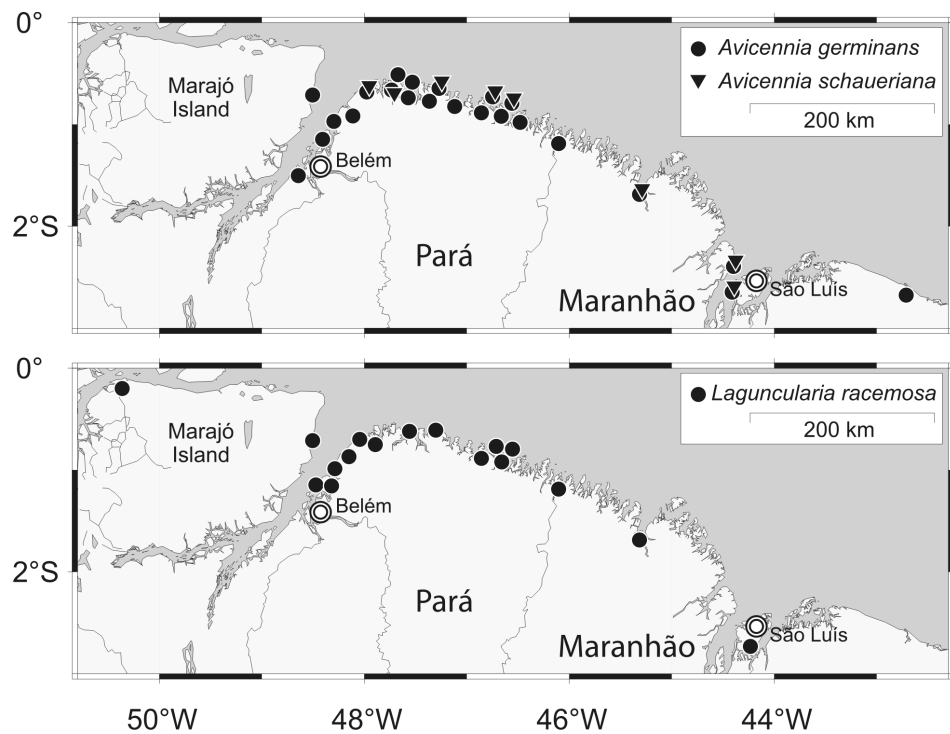


Figure 3 - Distribution of *Avicennia* species (above) and *L. racemosa* (below) on the study area. For details about the localities name and position, see Annex 2 and 3.

Minor elements of mangrove forest and associated species (*sensu* Tomlinson, 1986)

Among the species cited as mangrove associates (Table 1), only a few occur under high salinity conditions (salinity >30). Most of these are halophytic herbs more frequently found at channel borders or in other open areas; a few are salt marsh species (Carlton, 1975). The only woody high salinity mangrove associate is *Conocarpus erectus* L. (Combretaceae), which mainly colonises ecotones of mangroves and salt marshes where the soil is sandier and less frequently inundated. A few shrubs and vines (*Muellera*-Fabaceae-Faboideae, *Rhabdadenia* - Apocynaceae) have been observed by the authors in saline environments (salinity >30) on Ajuruteua Peninsula, Bragança. However, most of the woody mangrove associates, as well as herbs like *Crinum* sp. (Amaryllidaceae) and the mangrove fern *Acrostichum* (Pteridaceae), seem to be unable to thrive without significant freshwater influence. A number of species cited in the literature might in fact be of terrestrial origin (e.g., *Astrocaryum*, *Cassipourea*) and thus belong to adjacent ecosystems like *restinga* (dune vegetation). Other species may be occasional invaders from temporary flooded upstream fresh water habitats (Várzea).

FOREST STRUCTURE

Schaeffer-Novelli *et al.* (1990) divide the Brazilian coast into 7 segments. The mangroves of our study area ("segment III") are classified as *Rhizophora*-dominated "fringe forests", reaching 20 m in height. Nevertheless, our review reveals considerable heterogeneity in species composition and forest structure and indicates that this classification is too coarse. Unfortunately, the majority of studies is merely descriptive and does not provide quantitative data based on an adequate sampling area. Furthermore, the methods used to describe forest structure include a variety of plot-based methods and distance-based methods (point-centred quarter method), as well as "rapid assessment" estimation techniques. This means that comparisons of the results of the studies (summarised in Table 2) should be viewed with caution. A maximum average canopy height of approximately 25 m was observed by Ferreira (1989) in a mature mangrove forest in Vigia, Pará (location 9 in Figure 1). Average trunk diameter at breast height (dbh) can reach 42 cm in Curuçá, location 12 in Figure 1 (Menezes, 1994). Individual *A. germinans* trees occasionally attain 100 cm dbh in Salvaterra, location 4 in Figure 1 (Lisboa *et al.*, 1993); even larger trees with heights over 30 m and diameters greater than 1.5 m were observed by Thüllen (1997) at Furo do Meio, on Ajuruteua Peninsula (location 21 in Figure 1).

In hypersaline and rarely inundated areas, *A. germinans* forms monospecific dwarf forests (Santos *et al.*, 1997; Reise, 2003) with tree heights of less than 2.5 m (Table 2). Lara and Cohen (2006) observed an inverse relationship

between salinity and tree height. Studies of leaf gas exchange (Brabo, 2004) and chlorophyll fluorescence (Hu, 2004) of *A. germinans*, *A. schaueriana* and *L. racemosa* trees indicate that salt stress reaches critical levels under conditions found in a rarely inundated, hypersaline (salinity ~90) dwarf tree stand on Ajuruteua Peninsula. However, no comparable stress response was detected at a moderately saline, well-inundated site (Brabo, 2004; Hu, 2004). Likewise, results of dendrochronological analyses undertaken by Menezes *et al.* (2003a) in the same area show similar stem growth rates of *R. mangle* trees from brackish and from saline, well inundated mangrove stands. However, additional research is required to improve our understanding of the influence of environmental factors on the growth of Amazonian mangrove trees and, consequently, on the structure of the mangrove forests in the region. The scope of current research should be broadened to include other tree species and cover additional study sites besides Ajuruteua Peninsula.

Satellite image analyses carried out for the Ajuruteua Peninsula reveal geobotanical units (Souza-Filho and Paradella, 2002; Cohen *et al.*, 2004) which can be mapped to the forest types identified by terrestrial forest structure analysis (Thüllen, 1997, Reise, 1999; 2003; Mehlig, 2001; Menezes *et al.*, 2003a). By matching the results obtained by both the remote sensing and the ground-based approach (and from analyses of flooding regimes and salinity conditions undertaken by Cohen *et al.*, 2004), the following mangrove forest types can be distinguished on Ajuruteua Peninsula:

1) *Avicennia germinans* dwarf forest (mean tree height <2.5 m; Reise, 1999; 2003) at highest elevations (above mean spring tide level, inundation frequency <28 d.y⁻¹); pore water salinity 90–100.

2) Low *A. germinans* forest (mean tree height 6 m, mean dbh 5 cm; Reise, 1999) in areas inundated only during normal spring tides (28–78 d.y⁻¹); pore water salinity 90–50.

3) *A. germinans*-dominated, tall mixed forests (mean height 11.8 m, mean dbh 19.2 cm; Menezes *et al.*, 2003a); inundated for about 175 d.y⁻¹; pore water salinity 23–58.

4) Tall, mixed *R. mangle*/*A. germinans* forests (mean height 12 m, mean dbh 31 cm; Menezes *et al.*, 2003a), composed of *R. mangle*, *A. germinans* and *Laguncularia racemosa*, but dominated by *R. mangle*; at mid-tide level, normally well inundated (up to 233 d.y⁻¹); pore water salinity 20–60.

4a) A variety of the forest type Tall, mixed *R. mangle*/*A. germinans* forests (type 4) occurs under brackish conditions; it has similar structural properties, but features a greater number of associated understorey species (Menezes *et al.*, 2003a; Berger *et al.*, 2006). The pore water salinity ranges from 10 to 17 and inundation frequency is about 140

Table 3- Phenology of Amazonian mangrove species. ● = new leaves/flowers/fruits observed; ■ = peak period.

Species	J	F	M	A	M	J	J	A	S	O	N	D
<i>A. germinans</i> ^{1,2,5}												
leaves	●	●	●	●	●	●	●	●	■	■	■	●
flowers	●							●	■	●	●	●
fruits	●	●	■	■								●
<i>A. schaueriana</i> ^{3,5}												
bud						●	●					
flowers				●	●	■	■	●	●	●		
fruits	●	■	■	●		●	●	●				●
<i>R. mangle</i> ^{1,2,4,7}												
leaves	■	■	■	■	■	■	●	●	●	●	●	●
flowers	●	●	●	●	■	■	■	■	●	●	●	●
fruits	●	●	●	●	●	●			■	■	■	●
propagules	■	■	■	●								
<i>R. racemosa</i> ³												
buds		●	●		●	●			●	●		●
flowers	?											
fruits	●	●	●	●	●	●	●	●	●	●	●	●
propagules			●	●					●	●		●
<i>L. racemosa</i> ^{1,6}												
leaves	●	■	●	●	●	■	●	●	●	●	●	●
flowers	●	●	●	●	●							●
fruits	■	■	■	■	●							

References: ¹Mehlig (2001); ²Carvalho (2002); ³MPEG Herbarium (MG); ⁴Menezes (1997); ⁵Santos (2005); ⁶Silva (2005); ⁷Mehlig (2006)

d·y⁻¹. These forests occupy inner parts of the estuary with pronounced influence of fresh water runoff.

In addition to the forest types described so far, other forest types can be identified as: 1) regenerating forests consisting of seedlings and/or young trees of *L. racemosa* and *A. germinans*; 2) degraded areas with totally or partially dead vegetation (due to damage caused by human activities or changes in local conditions).

All types of forests occur side by side on the peninsula as can be seen clearly in aerial photographs (Reise, 2003) and satellite images (Souza-Filho and Paradella, 2002; Cohen *et al.*, 2003). The forest types so far described indicate that in spite of the low species richness there is a great deal of variation in mangrove forest structure, corresponding to topographical and hydrological conditions. Since the geomorphological settings and hydrographic conditions in other regions of the Brazilian north coast are comparable to those on Ajuruteua Peninsula (Souza-Filho and Paradella, 2002), vegetation patterns observed here are expected to be typical of the region as whole. Investigations on forest structure of other areas should confirm this conclusion

PHENOLOGY OF MANGROVE TREES

Information on mangrove phenology is derived from studies of litter fall (Mehlig, 2001; Reise, 2003; Mehlig, 2006; all from Bragança - locality 21 in Figure1), direct observation of phenophases of whole trees (Carvalho, 2002 in Bragança - locality 21 in Figure1), and periodic monitoring of individually marked shoots (Menezes, 1997 at Curuçá, locality 12 in Figure 1; Silva, 2005; Santos, 2005; Mehlig, 2006 - all at Bragança - locality 21 in Figure1). Information on the phenology of *R. racemosa* is based on an analysis of specimens from the João Murça Pires Herbarium at the Paraense Emílio Goeldi Museum fo Pará (MG) in Belém. There are no data on phenology of *R. harrisonii*.

A. germinans leaf fall is seasonal and precedes flowering (Mehlig, 2001; Carvalho, 2002). New leaves appear mainly during the transition from rainy to dry season when the salt stress is lower (Mehlig, 2001; Santos, 2005). Median leaf life time is 275 days (Santos, 2005). Flowering is restricted to the dry season (Mehlig, 2001; Carvalho, 2002; Santos, 2005). Ripe fruits are released in the rainy season (Table 3; Carvalho, 2002). The time span between peak of flowering and fruit fall is 4-5 months (Mehlig, 2001).

Information on the phenology of *A. schaueriana* is based on observations by Santos (2005). Leaves are produced throughout the year but higher production rates were observed during the rainy season. No clear pattern in leaf

Table 4 - Annual litter fall rates (Mg⁻¹·ha⁻¹·y) from different locations on Ajuruteua Peninsula, Bragança, Pará state.

Site	Position	Litter dry matter (Mg ⁻¹ ·ha ⁻¹ ·y)	Reference
Furo do Meio	0°52'17"S;46°39'00"W	12.8	Mehlig (2001)
Acarajó	1°00'44"S;46°45'44"W	13.1	Mehlig (2001)
Furo Grande I	0°50'19"S;46°38'15"W	7.9	Carvalho (2002)
Furo Grande II	0°50'47"S;46°38'51"W	15.4	Reise (2003)
Dwarf Forest	0°53'98"S;46°40'32"W	4.1	Reise (2003)
Furo Grande III	0°50'25"S;46°38'20"W	16.3	Nordhaus (2004)
Central peninsula	0°55'40"S;46°40'20"W	11.8	Nascimento <i>et al.</i> (2006)
Bosque de <i>Avicennia</i>	0°55'65"S;46°40'09"W	9.5	Batista (2003)

fall is discerned. The median life span of leaves is 115 days. Flowering takes place from April to October; ripe fruits are observed during the rainy season (Table 3).

L. racemosa sheds leaves throughout the year (Mehlig, 2001). New leaves are produced throughout the year at well-inundated sites but production rates are higher during the rainy season. Median leaf life time is 100 days (Silva, 2005). *L. racemosa* starts flowering mostly at the beginning of the rainy season (Table 3). Fruits develop quickly and are shed in the rainy season (Mehlig, 2001; Silva, 2005).

R. mangle sheds leaves throughout the year and new leaves are also produced continuously. Leaf production is higher in the rainy season (Table 3), when soil salinity is lower. Leaf age is variable but rarely exceeds one year (Mehlig, 2006). *R. mangle* also produces flowers throughout the year, but flowering peaks at the end of the rainy season (Menezes, 1997; Mehlig, 2001; Carvalho, 2002; Mehlig, 2006). Propagule release is mostly restricted to the rainy season (Menezes, 1997; Mehlig, 2006; Carvalho, 2002). The time span between pollination and maturation of propagules is about 8 months (Menezes, 1997; Mehlig, 2006). According to Mehlig (2001), the comparison of propagule production between brackish and saline mangroves suggests that differences in salinity regime might play a role in premature abscission of flowers and fruits.

Buds and immature propagules are present in almost all examined herbarium specimens of *R. racemosa*. It is thus probable that flowering and propagule development in *R. racemosa* take place throughout the year (Table 3).

Generalisations based on the currently available phenological data should be avoided due to the small geographic range covered so far. However, differences in patterns of rainfall (for example in the duration of rainy and dry seasons) probably influence the extension and intensity of the phenophases.

LITTER FALL

Information about mangrove litter fall is available from Ajuruteua Peninsula (Schories and Reise, 2000; Mehlig, 2001; Carvalho, 2002; Reise, 2003; Nordhaus, 2004; Table 4). In all studies, leaves are the most abundant component of the litter (about 70% of total litter). Litter fall rates range between $4.1 \text{ Mg}^{-1} \cdot \text{ha} \cdot \text{y}^{-1}$ in *A. germinans* dwarf forest (Schories and Reise, 2000) and $16.3 \text{ Mg}^{-1} \cdot \text{ha} \cdot \text{y}^{-1}$ in tall, mixed *R. mangle*/*A. germinans* forest (Nordhaus, 2004; Nordhaus *et al.*, 2006). Comparing mangroves with similar structural characteristics on Ajuruteua Peninsula, Schories and Reise (2000) identify salinity as the principal factor affecting the total litter production; the most productive forests are frequently inundated and are characterised by lower salinity. Schories *et al.* (2003) estimate that less than 31% of the litter

production of a mangrove stand at Furo do Meio on Ajuruteua Peninsula is exported to the estuary by tidal flushing.

LEAF DECOMPOSITION

Leaf decomposition was analysed through mesh bag exposure experiments by Sales (2000), at São Caetano de Odivelas east of the Marajó Bay ($0^{\circ}44'50''\text{S}$, $48^{\circ}01'16''\text{W}$), and Schories *et al.* (2003) at Furo do Meio on Ajuruteua Peninsula. In both studies, *A. germinans* leaves decomposed faster than those of *R. mangle* and *L. racemosa*. Decomposition of 50% of the exposed material lasted 34 and 90 days for *A. germinans* and *R. mangle*, respectively (Sales, 2000). Schories *et al.* (2003) found slightly faster decomposition rates at frequently inundated than at rarely inundated sites (23-34 vs. 29-41 days, respectively). Sales (2000) unfortunately does not specify the inundation regime at his experimental site.

HERBIVORY

A. germinans is described as the tree species most affected by herbivores, while *R. mangle* is considered the least affected species (Pontes and Mochel, 2000; Rebelo-Mochel, 1997; Ohana *et al.*, 1996; Praxedes and Mello, 1998; Praxedes *et al.*, 1998). The most common predators of mangrove vegetation are Hymenoptera (Praxedes and Mello, 1998; Pontes and Mochel, 2000), crustaceans, Isoptera, Hemiptera (Pontes and Mochel, 2000) and Lepidoptera (Pontes and Mochel, 2000; Menezes and Mehlig, 2005). According to Pontes and Mochel (2000), *L. racemosa* is severely affected by herbivores in areas polluted by waste-water discharges. Pontes and Mochel (2000) attribute herbivore activity on *L. racemosa* trees to increased nitrogen concentration in the leaves caused by sewage. A massive infestation of *A. germinans* by caterpillars of the moth *Hyblaea puera* (Hyblaeidae) was documented on Ajuruteua Peninsula in 1998 (Mehlig and Menezes, 2005), leading to complete defoliation of *A. germinans*-dominated stands. This phenomenon contributed to a decrease in litter fall (Mehlig, 2001). Similar observations were made in the same area in 2000 (Reise, 2003) and in São Caetano de Odivelas (Sousa, 2001; Prost *et al.*, 2001).

ANTHROPIC USES OF MANGROVE FLORA

Mangrove wood is utilized as fuel and as material for rural construction; poles are extracted for the construction of fish traps (Bastos, 1995; Almeida, 1996b; Rebelo-Mochel, 1997; Glaser and Grasso, 1998; Santana, 2001; Senna *et al.*, 2002; Glaser *et al.*, 2003). *Rhizophora* bark is used for tannin extraction. Glaser *et al.* (2003) identify two types of timber extraction: "basic needs" extraction and extraction for commercial purposes. "Basic needs" extractors utilize mangrove products for subsistence (domestic fuel, fish traps). Commercial extraction of wood and bark is an activity not talked about openly because it is illegal under Brazilian law.

However, the commercial use of the mangrove trees contributes to the income of a considerable number of households (Glaser *et al.*, 2003). *R. mangle* is the most exploited species. Almeida (1996b) explains the apparent preference for *R. mangle* wood over that of the *Avicennia* species as reflecting the greater abundance of the former species, but considers wood of *Avicennia* as equally suitable or even better for construction. Large *Rhizophora* trees can be felled quite easily by cutting their stilt roots. Large *A. germinans* trees are often hollow and thus not suitable for certain uses. However, large trees of either species are rarely extracted because the transport of logs through the mangrove swamp is usually not viable (they do not float) and stems cannot always be sawn into transportable pieces *in situ* (pers. obs.).

Bark, leaves and young roots from *R. mangle*, *Conocarpus erectus* and the grass *Spartina alterniflora* are used in traditional popular medicine (Bastos, 1995; Almeida, 1996b). During its flowering period, *A. germinans* attracts honeybees kept by locals. "Mangrove honey" is an appreciated product in the region, but its commercial potential has not been assessed so far. To collect the honey of feral honey bees, locals occasionally cut down trees (Rebello-Mochel, 1997).

IMPACTS OF HUMAN ACTIVITIES AND PERSPECTIVES OF MANGROVE RECOVERY

The coastal zone is subjected to diverse anthropogenic pressures due to exploitation and human occupation (Szlafsztein, 2003). Mangrove areas are impacted by clay extraction, aquaculture, agriculture, salt extraction and industrial activities (Senna *et al.*, 2002), and may also be affected by recreational activities (Szlafsztein, 2003). Near urban centers, pollution can affect the ecosystem; Rebello-Mochel *et al.* (2001) describe oil spillages and waste dumping as responsible for mangrove destruction around the city of São Luis. On the other hand, mangroves are important for subsistence of the coastal population. For example, in coastal villages of Bragança district, 80% of the population depend on mangrove products for a variety of subsistence uses (Glaser, 2003). Tourinho (1998) reports the establishment and subsequent abandonment of rice paddy fields in freshwater-influenced mangroves near Acarajó in the Caeté estuary. Rice cultivation failed because farmers were unable to control invasive weeds. Secondary succession of abandoned areas started with colonisation by salt marsh and swamp vegetation, then by *L. racemosa* and finally by *A. germinans*. However, regenerating areas are still distinctly different from the original *R. mangle* dominated forest, which is preserved in the surrounding areas. This case study shows that, even under benign environmental conditions, the recovery of deforested or degraded areas to the original ecosystem state is not straightforward (Berger *et al.* 2006).

Changes of the hydrographic regime caused by construction of a causeway through mangrove forests on Ajuruteua Peninsula have led to die-back of more than 2 km² of mangrove forest along a 25 km section of the PA 458 highway. The new road blocked several tidal channels, reducing the inundation frequency in some areas and leading to prolonged or permanent inundation in others (Souza-Filho, 2001; Carvalho, 2000; Krause *et al.* 2001). A natural resilience is observed in parts of the area, where *L. racemosa* trees are successfully colonising the bare mudflat. According to Cohen and Lara (2003), the degraded area decreased by 6.6% per year during 1986-1988 and 0.04% in 1991-1997.

Loss and gain of mangrove vegetation along the coast line of Pará has been monitored by means of satellite images (Souza-Filho, 2001; Souza-Filho and Paradella, 2002; Cohen and Lara, 2003). In 1972 the same study area had a mangrove coverage of about 592 km², which declined to 585, 583 and 573 in 1986, 1991 and 1997, respectively. The overall net loss of mangrove vegetation coverage in the 1972-1997 period was 19 km², a total net loss of 3.2%, which is approximately 0.76 km²·yr⁻¹ (Cohen and Lara, 2003). Mangrove losses are associated with erosion or are the result of sand deposits asphyxiating mangrove roots (Souza-Filho, 2001; Souza-Filho and Paradella, 2002; Cohen and Lara, 2003). New mangroves can develop following the colonisation of emerging mud banks by *Spartina* grass or *Laguncularia racemosa* (Prost *et al.*, 2001; Menezes *et al.*, 2003b).

Analyses of Landsat images and pollen profiles from soil cores by Behling *et al.* (2002) document a long-term expansion of mangroves on Ajuruteua Peninsula towards higher elevation salt flats with herbaceous vegetation. This expansion may be due to rising sea levels (Lara and Cohen, 2002).

CONCLUSION AND OUTLOOK

This review demonstrates that the knowledge about mangrove vegetation in Amazonia is still somewhat scarce. The majority of studies is based on rapid assessment making a more detailed analysis difficult. A notable exception are the studies on Ajuruteua Peninsula, Bragança, Pará. However, it seems necessary to replicate the research effort made at this particular point in a number of other places along the Brazilian north coast before being able to describe the mangroves of Amazonia more conclusively. Further, the almost complete lack of long-time monitoring studies e. g. of litter fall or of forest development makes it difficult to elaborate management decisions regarding mangrove vegetation. The recent creation of Marine Extractivist Reserves (*reservas extrativistas marinhas*, RESEXM) by the Brazilian Institute for Environment and Renewable Natural Resources (*Instituto Brasileiro do Meio Ambiente e Recursos Naturais Renováveis*, IBAMA) along the coast of Pará state documents that decision

makers have recognized the need to protect the important resources provided by mangrove vegetation. However, without supplementing the current knowledge through continuation and intensification of mangrove research, the success of mangrove management is at stake.

Nevertheless, the information available so far already allows to characterise the north Brazilian mangroves with more detail than has been done so far, and permits to formulate more specific future research goals. Previous works have underlined chiefly the supposed climatic optimum for mangrove growth near the equator, leading to “better developed” mangrove forests in Amazonia than in the Brazilian Northeast and South (Schaeffer-Novelli *et al.*, 1990; Kjerve and Lacerda, 1992; Lacerda, 1999). Saenger and Snedaker (1993) expect that structural complexity of mangroves is increasing towards the northern and southern geographical limits of mangrove vegetation. While indeed mangrove forests under moderate salinities and high inundation frequencies reach greater tree heights and stem diameters along the north coast than further south, a closer look reveals that climatically benign conditions cannot compensate for growth limitations imposed by other factors, in particular low inundation frequency and the concomitant increase in soil pore water salinity. Consequently, structural variability and floristic differences within mangroves of the Brazilian north coast are greater than reported in the past. Likewise, litter fall rates are not homogeneously high but vary between <4 and $>16 \text{ Mg}\cdot\text{ha}^{-1}\cdot\text{y}^{-1}$, corresponding closely to structural development of the forest stand in question. While maximum litter fall rates are world-wide among the highest reported so far as predicted by Saenger and Snedaker (1993) for low-latitude mangroves, a considerable portion of Brazilian north coast mangroves produces but a fraction of this maximum. Therefore, litter fall-based primary productivity estimates will require careful evaluation of the contribution made by the different forest types and their characteristic litter fall volume.

Interestingly, all mangrove tree species within our study area show certain phenological patterns. As the temperature regime of coastal areas near the equator is comparatively even, rain fall seasonality with its marked influence on the salinity of coastal water bodies is the most promising trigger for initiation/termination of phenophases. Observing correlations of rain fall/salinity with water balance, reserve substance accumulation and phytohormon levels during the course of the year would shed light on the underlying physiological processes; however, such studies have not yet been conducted anywhere in mangroves.

The distribution of two of the species of the genus *Rhizophora*, *R. racemosa* and *R. harrisonii*, is disjunct with distribution centres in the Marajó Bay (Pará state) and the River Preguiças estuary. Mangroves of Marajó Bay (Pará state)

are strongly influenced by freshwater through the discharge of the River Tocantins, River Pará and other rivers leading to the Bay. In contrast, the neighbouring section of the coast of Pará is characterised by higher salinity levels. Inability of these species to cope with higher salinity, or to compete with congeneric *R. mangle* and other mangrove trees under high salinity conditions is a suggestive explanation. However, salinity data are not available for the Preguiças estuary, neither are data on possible genetic isolation of the two populations which would provide hints for an explanation of the biogeographic history of these *Rhizophora* species in Brazil. In addition, the assumed physiological differences between *R. racemosa* and *R. harrisonii* on one hand and *R. mangle* on the other still have to be confirmed experimentally, as mentioned before.

In contrast to mangroves in other regions of Brazil (e. g., Soares, 1999), Amazonian mangroves are well preserved, possibly due to the still less severe population pressure in the coastal regions of the study area, and due to the more difficult access (Lara, 2003). Nevertheless, mangroves are endangered by population growth, deforestation for a variety of purposes, among them the increasing demand for wood (Senna *et al.*, 2002; Szlafsztein, 2003). Policy initiatives are required for the effective preservation of mangrove forests along the Brazilian north coast. The before-mentioned approach to create RESEXM co-management reserves (Glaser and Oliveira, 2004) focuses mainly on sustainable use of resources important for the human population and on preserving traditional forms of resource extraction. To complement this, we emphasise the importance of current and future research. Continued investigation, focusing on topics such as species distribution, plant growth, adaptation to environmental factors and plant associations, is essential for the preservation of mangroves in the region.

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Annex 1 - Occurrence of *Rhizophora* species along the coast of Pará and Maranhão states. MG indicates data from the herbarium of MPEG .

Local	Rm	Rr	Rh	References	Position
Pará State					
Afuá	•			Maciel and col. 1842 (MG 144453)	0°09'56''S;50°24'8''W
Ananindeua		•		Ferreira (1989)	1°22'12''S;48°22'42''W
Augusto Corrêa: Urumajó	•			Carreira and col. 1462 (MG 165749), Bastos <i>et al.</i> (2000); Bastos <i>et al.</i> (2002)	1°01'12''S;46°39'33''W
Barcarena: Rio Caripi, Vila São Vicente		•		Lins and col. 338 (MG 115416)	1°30'60''S;48°37'53''W
Belém: Ilha das Onças, Ilha da Tatuoca	•	•		Prance <i>et al.</i> (1975); Almeida and Lobato 376 (MG 135338)	1°26'22''S;48°33'47''W
Benevides: Benfíca		•		Rosa and Renner (MG 0147650); Ferreira (1989)	1°21'49''S;48°15'7''W
Bragança: Península de Ajuruteua, Ilha de Canelas	•			Almeida (1996b); Thüllen and Berger (2000); Krause <i>et al.</i> (2001); Amaral <i>et al.</i> (2001); Menezes <i>et al.</i> (2003); Matni, (2003); Seixas (2003); Abreu (2004)	1°03'06''S;46°45'44''W
Colares		•		Almeida (1996a)	0°55'59''S;48°17'12''W
Curuçá	•			Mello <i>et al.</i> (1995); Carvalho <i>et al.</i> (1995); Almeida (1996a)	0°44'5''S;47°52'20''W
Ilha de Algodoal; Praia do Crispim; Ilha de Maiandea: Marudá	•			Bastos and Lobato (1996), Carreira <i>et al.</i> (2002)	0°37'5''S;47°34'14''W
Ilha de Mosqueiro: Praia do Maraú, Praia do Paraíso		•		Prance <i>et al.</i> (1975)	1°8'1''S;48°27'29''W
Marajó: Breves, Salvaterra, Soure, São Sebastião da Boa Vista	•	•		Lisboa <i>et al.</i> (1993); Rosa and col. 5469 (MG 143964); Rosa and col. 5547 (MG 144042)	0°41'53''S;48°31'15''W
Marituba	•	•		Almeida (1996a)	1°21'20''S;48°20'14''W
Marapanim: Rio Marapanim, Rio Mojuim, Rio Cajutuba	•			Prost and Loubry (2000); Costa-Neto <i>et al.</i> (2000); Prost <i>et al.</i> (2001); Sales (2005); Berredo (2005)	0°42'57''S;47°41'43''W
Primavera,	•			Almeida (1996a)	0°56'29''S;47°06'58''W
Quatipuru	•			Huber 1899 (MG 1766); Almeida <i>et al.</i> (2002)	
Salinas: Rio Sampaio	•	•	•	Prance <i>et al.</i> (1975); Ferreira (1989)	0°37'31''S;47°21'43''W
Santa Bárbara		•		Santos <i>et al.</i> 2 (MG 172748)	1°13'16''S;48°17'36''W
São Caetano de Odivelas	•			Prance <i>et al.</i> (1975); Ferreira (1989); Sales (2000) Prost and Loubry (2000); Luz <i>et al.</i> (2000);	0°44'49''S;48°1'17''W
São João de Pirabas	•			Menezes <i>et al.</i> (2003)	0°46'05''S;47°10'44''W
Vigia	•	•		Ferreira (1989); pers. obs.	0°51'40''S;48°08'35''W
Viseu: Jabotitua-Jatium	•			Santos <i>et al.</i> (2003) Menezes <i>et al.</i> (2003b)	1°12'18.2''S;46°08'18.8''W
Maranhão State					
Baía de São Marcos	•			Silveira and Mochel (2000)	2°40'38.68''S;44°31'31.8''W
Golfão Maranhense	•	•	•	Santos (1986)	2°45'43.43''S;44°21'69''W
Rio Mearim	•			Behling and Costa (1997)	
Rio Preguiças	•	•	•	Santos (1986)	2°37'06.23''S;42°41'41.28''W
São Luis	•	•		Prance <i>et al.</i> (1975); Damasio (1980a); Rebelo-Mochel (1997); Rebelo-Mochel <i>et al.</i> (2001);	2°31'03''S;44°17'59''W
Turiçu	•			Rebelo-Mochel (2000)	1°40'15''S;45°22'01''W

Annex 2 - Occurrence of *Avicennia* species along the coast of Pará and Maranhão states. MG indicates data from the herbarium of MPEG .

Local	Ag	As	Reference	Position
Pará State				
Ananindeua: Maguarizinho, Rio Seco, Curuçambá	•		Ferreira (1989)	1°22'12"S;48°22'42"W
Augusto Correa: Urumajó	•		Carreira and col. 1449 (MG 165737); Bastos <i>et al.</i> (2002)	1°01'12"S;46°39'33"W
Barcarena	•		Lins and col. 103 (MG 113223)	1°30'60"S;48°37'53"W
Belém: Ilha das Onças	•		Huber 1903 (MG 3832)	1°26'22"S;48°33'47"W
Benevides: Igarapé Estaleiro	•		Ferreira (1989)	1°21'49"S;48°15'7"W
Bragança: Península de Ajuruteua, Ilha de Canelas, Praia de Ajuruteua	•	•	Almeida (1996a); Thüllen (1997); Amaral <i>et al.</i> (2001); Matni (2003); Seixas (2003); Brabo (2004); Abreu (2004); Santos (2005)	1°03'06"S;46°45'44"W
Curuçá	•	•	Menezes (1994); Carvalho <i>et al.</i> (1995); Almeida (1996a); Mello <i>et al.</i> (1995)	0°44'5"S;47°52'20"W
Colares	•		Almeida (1996a)	0°55'59"S;48°17'12"W
Igarapé do Japim	•		Duke, 1895 (MG)	no information
Ilha de Algodão: Maiandeuá, Maracanã,	•		Bastos and Lobato (1996); Carreira <i>et al.</i> (2002)	0°37'5"S;47°34'14"W
Ilha de Marajó: Pacorral, Salvaterra	•		Gress 1899, 1915 (MG); Lisboa <i>et al.</i> (1993);	0°41'53"S;48°31'15"W
Marapanim:Rio Marapanim, Rio Cajutuba	•		Prost and Loubry (2000); Costa-Neto <i>et al.</i> (2000); Sales (2005);Berredo (2005)	0°42'57"S;47°41'43"W
Marituba, Mosqueiro	•		Almeida (1996a); Ferreira (1989)	1°21'20"S;48°20'14"W
Marudá, Crispim	•		Bastos and Lobato (1996)	0°37'5"S;47°34'14"W
Primavera, Quatipuru, Boa Vista	•		Almeida (1996a), Rodrigues 5072 (MG 35584); Almeida <i>et al.</i> (2002)	0°56'29"S;47°06'58"W
Salinas: Rio Sampaio	•	•	Ferreira (1989)	0°37'31"S;47°21'43"W
São Caetano de Odivelas:Rio Araciteua, Ilha Nova, Rio Mojuim	•	•	Ferreira (1989); Luz <i>et al.</i> (2000); Prost and Loubry (2000); Prost <i>et al.</i> (2001)	0°44'49"S;48°1'17"W
São João de Pirabas	•	•	Sales <i>et al.</i> , 2004	0°46'05"S;47°10'44"W
Vigia	•		Ferreira (1989); Mello <i>et al.</i> , 1995	0°51'40"S;48°08'35"W
Viseu: Jabotitiua Jatium	•		Santos <i>et al.</i> (2003); Menezes <i>et al.</i> (2003b)	1°12'18.2"S, 46°08'18.8"W
Maranhão State				
Alcântara	•	•	Rosa and Silva 2509 (MG 132535); pers. obs	2°24'27"S;44°25'22"W
Baía de São Marcos	•	•	Silveira and Mochel (2000)	2°41'42"S;44°31'16"W
Parnaçu	•	•	Rebello-Mochel <i>et al.</i> , (2001)	2°37'27"S;44°21'41"W
Estreito de Coqueiros, São Luis	•		Santana (MG 69921)	2°31'03"S;44°17'59"W
São Luis	•		Fern and Rob 96 (MG 129408); Damasio (1980a); Rebello-Mochel (1997); Rebello-Mochel <i>et al.</i> (2001)	2°31'03"S;44°17'59"W

Annex 3 - Occurrence of *Laguncularia racemosa* along the coast of Pará and Maranhão states. MG indicates data from the herbarium of MPEG .

Local	Reference	Position
Pará State		
Ananindeua – Igarapé Curuçamba	Ferreira (1989); Mello <i>et al.</i> (1995);	1°22'12"S;48°22'42"W
Bragança	Almeida (1996a); Tourinho (1998); Amaral <i>et al.</i> (2001); Menezes <i>et al.</i> (2003); Menezes and Compton (2003)	1°03'06"S;46°45'44"W
Colares	Almeida (1996a)	0°55'59"S;48°17'12"W
Curuçá	Menezes (1994); Mello <i>et al.</i> (1995)	0°44'5"S;47°52'20"W
Ilha de Algodoal	Bastos and Lobato (1996); Carreira <i>et al.</i> (2002)	0°37'5"S;47°34'14"W
Ilha de Mosqueiro	Almeida (1996a)	1°8'1"S;48°27'29"W
Marapanim	Luz <i>et al.</i> (2000); Costa-Neto <i>et al.</i> (2000); Sales (2005)	0°42'57"S;47°41'43"W
Marituba	Almeida (1996a)	1°21'20"S;48°20'14"W
Primavera	Almeida (1996a)	0°56'29"S;47°06'58"W
Salinas: Rio Sampaio	Ferreira (1989); Mello <i>et al.</i> (1995)	0°37'31"S;47°21'43"W
São Caetano de Odivelas: Rio Araciteua	Ferreira (1989); Mello <i>et al.</i> (1995); Prost <i>et al.</i> (2001)	0°44'49"S;48°1'17"W
Vigia	Pers. obs.	0°51'40"S;48°08'35"W
Viseu: Jabotitiua-Jatium	Santos <i>et al.</i> (2003); Menezes <i>et al.</i> (2003b)	1°12'18.2"S, 46°08'18.8"W
Maranhão State		
Parnauçu	Rebello-Mochel (2000)	2°37'27"S;44°21'41"W
São Luis	Damasio (1980a); Rebello-Mochel (1997); Rebello-Mochel (2000); Rebello-Mochel <i>et al.</i> (2001)	2°31'03"S;44°17'59"W

